

Colloid Formation at Waste Plume Fronts

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Highly saline and caustic tank waste solutions containing radionuclides and toxic metals have leaked into sediments at U.S. Department of Energy (DOE) facilities such as the Hanford Site (Washington state). Colloid transport is frequently invoked to explain migration of radionuclides and metals in the subsurface. To understand colloid formation during interactions between highly reactive fluids and sediments and its impact on contaminant transport, we simulated tank waste solution (TWS) leakage processes in laboratory columns at ambient and elevated (70 °C) temperatures. We found that maximum formation of mobile colloids occurred at the plume fronts (hundreds to thousands times higher than within the plume bodies or during later leaching). Concentrations of suspended solids were as high as 3 mass %, and their particle sizes ranged from tens of nanometers to a few micrometers. Calcium carbonate is always one of the dominant phases of the plume front colloids, while the other phases varied with solution pH and temperature. During infiltration of the leaked high- Na^+ waste solution, rapid and completed Na^+ replacement of exchangeable Ca^{2+} and Mg^{2+} from the sediment caused accumulation of these divalent cations at the moving plume front. Precipitation of supersaturated $\text{Ca}^{2+}/\text{Mg}^{2+}$ -bearing minerals caused dramatic pH reduction at the plume front. In turn, the reduced pH caused precipitation of other minerals. This understanding can help predict the behavior of contaminant trace elements carried by the tank waste solutions and could not have been obtained through conventional batch studies.

Introduction

Leakage of radioactive wastes and toxic metals into the subsurface has been identified at various DOE facilities (1). At the Hanford Site in Washington state, about $2 \times 10^5 \text{ m}^3$ of high-level radioactive and hazardous wastes are stored in 177 underground tanks, including 149 single-shell tanks. Over time, 67 of the single-shell tanks developed leaks (2). Radionuclides and other contaminants have been found at elevated concentrations in the vadose zone and groundwater beneath the single-shell tank farms (3). The nature of

pathways and mechanisms responsible for contaminant transport are not yet fully known, but such understanding is needed for future risk assessment and remediation.

Besides carrying fission products, some of the tank waste solutions were very saline and caustic, containing high concentrations of sodium, nitrate, and aluminum (4–6). Heat released from radioactive decay caused tank waste solutions to boil, and temperatures in underlying sediments were sustained above ambient levels. Because of the extreme chemical–physical conditions, leakage of these tank solutions into the sediments resulted in dissolution of primary minerals and precipitation of secondary minerals. Formation of secondary colloidal phases could have become associated with contaminants from the tank liquids through coprecipitation and adsorption. Subsequent colloid transport might have contributed to the deep migration of contaminants (7, 8). Under the DOE's initiative on cleanup of the Hanford Site, recent studies have resulted in a variety of findings relevant to contaminant transport, including identifying the mechanisms for sorption of Cs^+ in sediments (9, 10), the formation of cancrinite as a secondary mineral phase (11, 12), the geochemical evolution of waste plumes (13, 14), the effect of cation exchange on Cs^+ sorption (15, 16), and the effect of colloids on Cs^+ transport (17).

In recent studies (13, 14), we showed that three distinct zones occur within active saline–caustic waste plumes; the silicate dissolution zone, neutralization zone, and plume front. The plume front is the moving boundary between invading tank waste solution and displaced uncontaminated pore waters. In this paper we report findings on colloid formation at the plume front obtained from our laboratory simulations of the tank waste leakage process.

Experimental Section

Sediment. Hanford Formation “coarse sand”, a glaciofluvial sediment collected from the 200 East Area, was used in our experiments. The Hanford Formation comprises about 85% of the vadose zone of the site. The 200 East and 200 West Areas are the locations housing high-level waste tanks at the Hanford Site. The major mineral components are presented in Table 1. The median grain size is $350 \mu\text{m}$, and average grain density is 2.77 Mg m^{-3} . The chemical composition of a water extract from this sediment (water to soil mass ratio = 1:1, at 21 °C for 24 h) is presented in Table 2.

Synthetic Tank Waste Solution. The highly saline–caustic tank waste solutions at Hanford were generated from past Pu and U processing, in which aluminum nitrate solutions were used to dissolve Pu and U from spent fuel. After extraction of actinides, the residual acidic solution was treated with NaOH to keep Al in solution. The waste solution was then pumped into underground tanks for storage. A typical fresh waste solution was a 4 M sodium–nitrate–nitrite–aluminate–hydroxide solution (4). Upon self-boiling while in subsurface waste tanks, the original solutions became more concentrated and formed salt cakes. Because different tanks leaked at different times and over different periods, the precise chemical compositions of waste solutions at the times of leaking are unknown. Agnew et al. (4) estimated compositions of tank supernatants based on their comprehensive survey (Table 3). The synthetic tank waste solution (TWS) used in this study was based on the directly measured composition of single-shell tank SX-111 (Table 3), which is believed to be fairly representative of leaked TWS from the SX tank farm. This synthetic TWS is within the middle range of Agnew's (4) estimated composition of tank supernatant solutions. Radionuclides were not included in our synthetic TWS, because

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TABLE 1. Mineral Composition of Hanford Coarse Sand^a

	%	main minerals	%	clay minerals	%
sand	93	quartz	55	illite	40
silt	6	potassium feldspar	25	smectite	20
clay	1	sodium feldspar	15	mixed illite/smectite	20
		calcium carbonate	1	kaolinite	20
		others	<4		

^a From ref 18.**TABLE 2. Chemical Composition of Water Extract from Hanford Coarse Sand^a**

element	Na	K	Ca	Mg	Cl ⁻	SO ₄ ²⁻	HCO ₃ ⁻
concn (mM)	1.85	0.31	0.53	0.18	0.75	0.76	1.3

^a Water to soil mass ratio = 1:1; pH 8.3.**TABLE 3. Chemical Composition of Tank Waste Solutions**

component	our synthetic TWS (M)	measured SX-111 (M)	Agnew's estimate (M)
Al	0.805	0.805	0.7–1.9
CO ₃ ²⁻	0.168	0.168	0.2–0.3
OH ⁻	1.47	1.47	3–8.5
NO ₃ ⁻	4.72	4.72	2.3–6.3
NO ₂ ⁻	0.84	0.84	1–2.9
SO ₄ ²⁻	0.06	0.06	0.1–0.3
Na ⁺	8.05	8.05	5–14
pH	14.1 (21 °C)		> 14

this study focused on the behavior of the major elements. All salts added to the synthetic TWS were completely dissolved within hours of stirring at room temperature, resulting in a solution density of 1.37 Mg m⁻³. The solution was filtered (0.45 μm) before use.

Column Method. The Teflon columns used to contain Hanford sediments were of two lengths, 50 and 250 mm, with an inner diameter of 38 mm. Both end caps of each column had polypropylene screens (105 μm mesh) to support the sediment and yet allow the transport of suspended particles. Moistened sediments (water 7% by mass) were packed into columns to an equivalent dry bulk density of 1.65 (±0.05) Mg m⁻³, to represent the initial partially saturated vadose zone (about 30% saturation). Pore volumes for the 50- and 250-mm columns were 22.9 and 115 mL, respectively. To avoid gravity-induced flow fingering, solutions were injected upward from the bottom of vertically oriented columns. Effluents were collected from the top of a column at room temperature into Teflon tubes by use of a fraction collector. The influent solutions were injected with a syringe pump at a pore-water velocity of 100 mm day⁻¹. Because the sediment underneath waste tanks experienced a range of elevated temperatures, experiments were conducted at elevated and ambient temperatures. For the high-temperature tests, columns were set inside a water bath, with the temperature controlled at 70 ± 0.5 °C. The ambient temperature tests were conducted at 21 ± 1 °C.

Two sets of column experiments were conducted, with 50- and 250-mm tall columns. In the 50-mm tall set, nine pore volumes (PVs) of TWS were injected and followed by nine PVs of the leaching solution to test infiltration-induced colloid release. The leaching solution was a 10-fold dilution of the soil water extract shown in Table 2. This solution was used to represent a mixture of native pore water from overlying uncontaminated soil and infiltrating rainfall water. Besides measuring pH and turbidity of the effluent fractions, precipitates contained in the plume-front effluents were isolated for bulk chemical analyses (electron microprobe)

and mineralogy identification (X-ray diffraction, XRD). The precipitates were collected on filters (0.25 μm), lightly rinsed with deionized water to remove excess soluble salts, and air-dried. In the 250-mm tall column set, only 6 PVs of TWS were injected. In addition to measuring turbidity and pH, concentrations of major cations were measured in these effluents.

Characterization of Effluent Solutions and Colloids. All measurements were conducted at room temperature. The pH measurements were obtained with a Corning semimicro combination glass pH electrode. This electrode was connected to a Denver Instrument model 225 pH/ISE meter. Calibrations were done with pH 4, 7, 10 (JT Baker), and 12.45 standards (Oakton). Recall that many of the experiments were conducted at high ionic strengths and elevated temperatures, whereas pH values were measured at room temperature. The reported pH values were intended for comparisons among the different systems rather than for inferring actual proton or hydroxyl activities under these extreme experimental conditions. Relative concentrations of suspended particles were measured by turbidity values [Hach model 2100AN turbidimeter, with a measurement range of 0–10 000 NTU (nephelometric turbidity units)]. The effluent fluids were ultrasonicated (Ultrasonic Cleaner, Cole Parmer 8892) for 10 min to disperse aggregates of particles before turbidity measurements were conducted. Particle size distributions in the plume-front effluent were measured on a centrifuge particle sizer (CPS; CPS Instrument, Inc.). Only the effluent fraction collected at the plume front contained enough particles for the particle size analyses.

Concentrations of major cations in effluents from the 250 mm tall columns were analyzed by inductively coupled plasma optical emission spectroscopy (ICP-OES; Thermo Jarell Ash Hi Res IRIS). Because many of these effluents contained colloids, and some of them formed after the effluents were collected, an aqua regia, hydrofluoric acid, and boric acid digestion was used to dissolve precipitates (19). The precipitates collected from effluent liquids were analyzed for their morphology (scanning electron microscopy, SEM), chemical compositions of individual particles (energy-dispersive X-ray microanalysis, EDX), mineralogy (XRD), and bulk chemical composition (electron microprobe). For the bulk chemical composition analyses, the air-dried precipitates were further dried at 110 °C, weighed, ignited at 925 °C, and weighed again to measure loss on ignition. Samples were fused with Li borate flux (1:1 Li₂B₄O₇:Li₂BO₃) in a Pt₉₅Au₅ crucible, in an oxidizing flame over an air–gas burner. Each sample was mixed with a Pt wire loop to make a bead, and the resulting bead was polished prior to determining major element concentrations by use of an electron microprobe (20, 21). Note that only the plume-front liquid contained enough mass of particles for the bulk chemical analyses.

Results and Discussion

Colloid Abundance in Effluents. The turbidity and pH breakthrough curves (BTC) under ambient (21 °C) and elevated (70 °C) conditions are presented in Figures 1 and 2a,b, for 50- and 250-mm tall columns, respectively. The turbidity and pH measurements were conducted as soon as effluent samples were collected. The vertical dashed line along the pore-volume axis in Figure 1a,b denotes the time of switching injection solution from the TWS to the leaching solution. Through these BTCs, we see the effluent began emerging from a column after about 0.6 PV of TWS injection, as expected on the basis of the calculated TWS displacement of the native sediment moisture (30% initial saturation). The first effluent fraction for the 50-mm tall and the first three fractions for the 250-mm tall columns contained primarily the displaced native soil water, which were characterized by

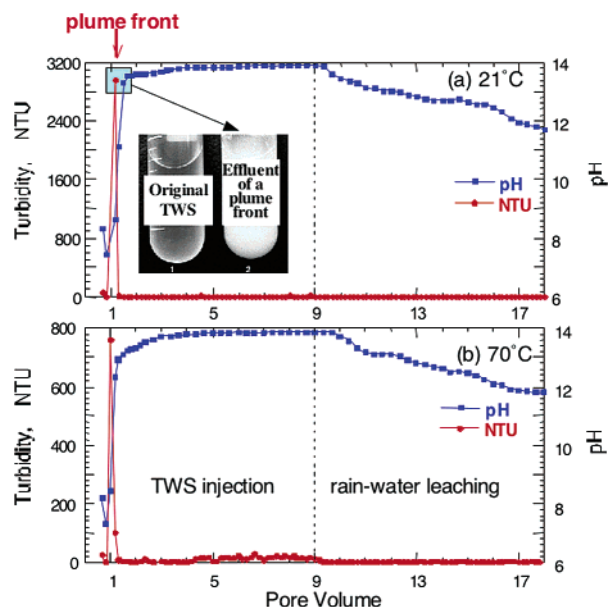


FIGURE 1. Turbidity and pH breakthrough curves from two 50-mm tall sediment columns at (a) 21 °C and (b) 70 °C. Nine PVs of the TWS were injected, followed by nine PVs of simulated rainwater. The vertical dashed lines along the x-axes indicate times for switching of these two solutions. The inserted photograph provides a comparison between the clear influent TWS and the milky plume-front effluent.

their pH values ≥ 8 (the initial soil water pH) and by their low salt concentrations (shown in Figure 2c–h). Also, the first native soil water fractions have relatively higher turbidity values than the second fractions (visible in Figure 1), because of suspended disturbed native soil particles. The second effluent fractions exhibit influence from mixing of the displaced soil water with TWS, indicated by its increased ionic strength, and decreased pH. The plume-front location can be determined by calculation of pore water displacement and also from the inflection point of the solution EC breakthrough, which indicates mixing of high-saline TWS with the displaced low-salinity native soil water. The phenomenon of dramatic pH reduction at the plume front has been reported recently (13, 14) and will be discussed further in this paper.

Figures 1 and 2b show that an extremely high and narrow turbidity peak occurred at the plume front (at about 1 PV), indicating formation and transport of large quantities of suspended particles. As shown in the Figure 1a inset, the plume-front particle concentration was so high that the effluent looked like milk. The peak turbidity values under ambient temperature conditions were always higher than that under 70 °C. The lower turbidities (less colloids) at 70 °C probably resulted from higher retention of Al within the column (see Figure 2h) and from larger retention of Si-associated solid phases. The highest turbidity sample (3400 NTU) obtained in the 21 °C effluents contained 3.0 mass % suspended solids. Particles were also detected within the plume body and the leaching stage. However, turbidity values of these later effluents were too low to be distinguished from zero on the linear axes used in Figures 1 and 2b. It should be noted that turbidity changed over time after collection of effluents within the plume body and the leaching stage. From the measurements over 60 days after the collection we found that turbidity generally increased, except in the plume-front fraction, where it decreased due to flocculation/aggregation. The increased turbidity over time is caused by newly formed colloids that are not the focus of this study. The greater extent of postcollection particle formation in the effluents from the 70 °C columns largely resulted from supersaturation with

respect to various minerals (discussed later) upon cooling to room temperature. Although the emphasis of this paper is not on these later precipitated particles, we determined their chemical compositions and structures in order to gain a more complete understanding of the reactions occurring in the plume.

Dramatic pH reduction of the reacted TWS was observed in the plume-front region (Figures 1 and 2a), where the pH decreased from 14 to as low as 7. Clearly, simple mixing of the TWS (pH 14) with native sediment water (pH 8) could not result in pH values < 8 . This suggests that the reactions resulting in colloid formation at the plume front are related to the large pH reduction (discussed further in a later section). Although both 21 and 70 °C columns exhibited similar pH decreases at the plume front, the pH of effluents collected after the front exhibited significant temperature dependence. The pH quickly increased back to 14 for the room-temperature condition, indicating that the reaction(s) causing pH reduction were slow relative to the supply of TWS. In contrast, the slower pH recovery under high-temperature conditions indicates that pH-reducing reactions remained significant. These different pH responses are indicative of strongly temperature-dependent reactions. Dissolution of silicates in caustic solutions is both strongly temperature-dependent and pH-neutralizing (22).

Colloids mobilization through leaching was expected because of the reacted sediment exposure to the much lower ionic strength solution. However, leaching did not mobilize significant amounts of colloids. Turbidity values in the leaching stage (measured immediately after an effluent sample was collected) were in the range of 0 to a few NTU (Figure 1). The lack of colloid mobilization during the leaching stage is probably related to the high ionic strength of the reacted leaching solution. The originally low ionic strength leaching solution became supersaturated with respect to various minerals (turbidity increased over time and mineral formation occurred over time *after* the effluents were collected). The original pH of the leaching solution was 7, but the effluent pH values ranged from close to 14 to 12 after eight pore volumes of leaching solution injection. The high ionic strengths suppress colloid dispersion. However, drainage of the leached column followed by rewetting with leaching solution might enhance displacement of colloids because advancing wetting fronts have been shown to be effective at colloid mobilization (23).

Chemical Composition of Effluents. Information on the chemical composition of effluents helps in understanding the plume-front processes and provides more perspective on how much transport is localized at the front. Major cation compositions of effluents from the 250-mm tall columns are presented in Figure 2c–h. These data are from analyses of acid-digested suspensions, such that concentrations represent sums of solution and suspended solid phases. Most notably, Ca^{2+} and Mg^{2+} were stripped off from cation exchange sites by the very high Na^+ concentrations in the invading waste solution (Figure 2c), leaving these divalent cations at extremely high concentrations within the front and practically undetectable later (Figure 2d,e). More Na^+ retardation at 70 °C resulted from formation and retention of Na–Si-bearing minerals (sodium cancrinite and sodium silicate) within the columns. The slightly higher retardation of Ca and Mg at 70 °C may also have resulted from precipitation of Ca/Mg-bearing phases. The K^+ curves have a similar cation-exchange peak localized at the plume front, followed by a broader period with elevated concentrations reflecting release from dissolution of soil minerals, especially at 70 °C. Although the localized Si peak at the plume front at 70 °C remains to be explained, Si removal (Figure 2g) resulted from temperature- and pH-dependent dissolution of silicate minerals in the sediment columns. Si became

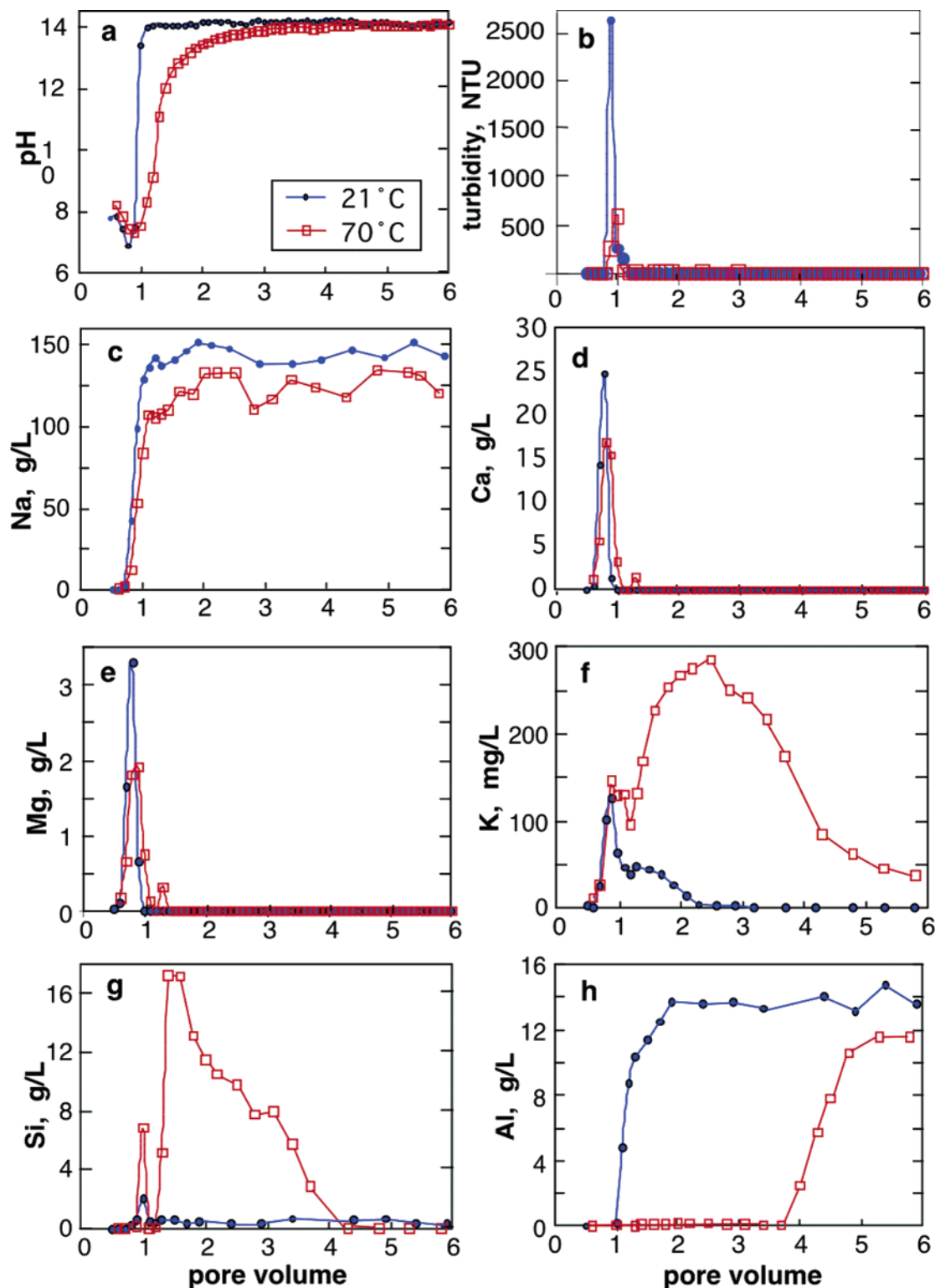


FIGURE 2. Breakthrough curves from two 250-mm tall columns (21 and 70 °C). These breakthrough curves of pH, turbidity, and the major cations were from six PVs of TWS injection.

nondetectable at the later time (after 4 PV), caused probably by coating-induced reduction of silicate dissolution surfaces and continuous precipitation of more stable Si-bearing solids such as cancrinites. The Al BTCs (Figure 2h) showed temperature and pH dependence. At the elevated temperature, reduced pH values (relative to pH 14) of effluents were also observed after emergence of the plume front (pH curves in Figure 2a), primarily because of silicate dissolution. Because of the reduced pH and dissolved Si, Al precipitated

and was retained within the column. After further permeation, the Si dissolution rate dropped and pH increased to nearly the initial pH of the TWS, thus allowing most of the Al originally in the TWS (0.80 M) to emerge in the effluent. At ambient temperature, most of the TWS Al emerged in the plume effluent solution because pH neutralization resulting from silicate dissolution was relatively weak. Aluminum precipitation at 21 °C occurred largely at the plume front where pH reduction was significant. It should be noted that

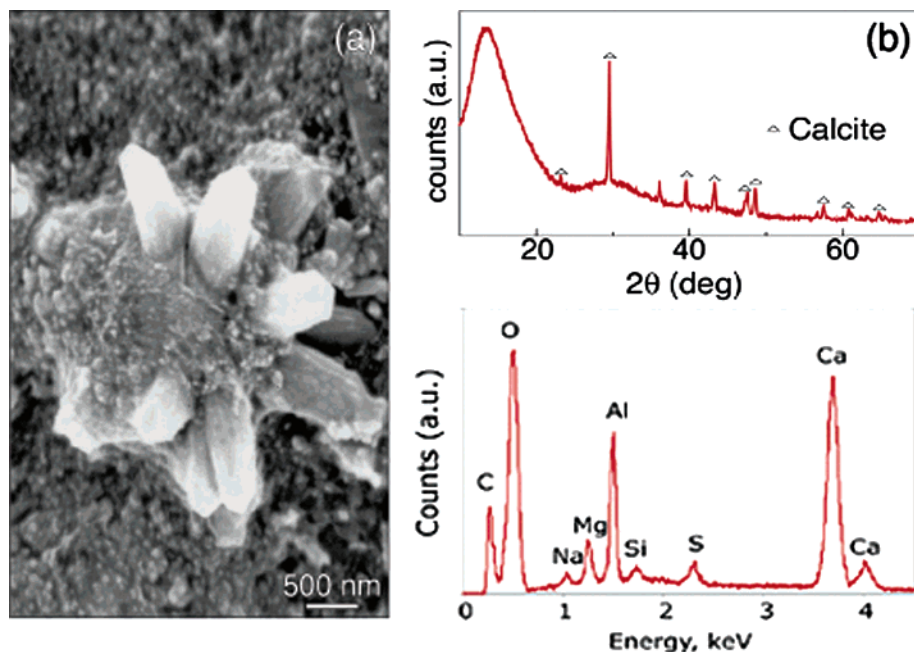


FIGURE 3. (a) SEM microphotograph of plume-front colloids at 21 °C, composed of fine amorphous particles and larger crystals. (b) XRD pattern of the bulk precipitates, indicating that calcite is the only crystalline phase. (Note: the large unmarked peak is the result of the sample holder used.) (c) Chemical composition from total area (10 μm^2) analysis.

TABLE 4. Chemical Compositions of Bulk Precipitates Collected from Plume-Front Effluents

temp	mass %						loss on ignition	Σ %
	SiO ₂	Al ₂ O ₃	Fe ₂ O ₃	MgO	CaO	Na ₂ O		
21 °C	0.34	18.18	0.04	5.38	26.88	0.46	45.90	97.18
70 °C	18.73	0.60	0.36	4.02	31.36	2.28	42.10	99.44

effluent Al concentrations never reached the initial concentration of the TWS of 0.80 M. This is because Al precipitated [Al(OH)₃ solubility is strongly pH-dependent] and was retained within the column.

Chemical Composition of Bulk Precipitates. Precipitates collected from the plume-front effluents of the 50-mm tall columns at two different temperatures were analyzed by electron microprobe. Concentrations of major cations and loss on ignition (primarily H₂O, OH⁻, CO₃²⁻, NO₃⁻, and SO₄²⁻) are presented in Table 4. Values of loss on ignition were 46 and 42 mass % for 21 and 70 °C systems, respectively. Calcium was the most abundant cation (27–31 mass % as CaO), and Mg concentrations were also high (5–4 mass % as MgO). Again Ca and Mg accumulated only within the narrow region of the moving plume front by cation exchange, displaced by Na (8 M) in the invading TWS, followed by precipitation of Ca/Mg solids. The next most abundant cations within the plume-front precipitates were Al at ambient temperature and Si at high temperature. Their precipitation at the plume front resulted from encountering reduced pH. These results are consistent with data on total elemental concentrations in effluents presented in the previous section.

Morphology and Mineralogy of Colloids. The morphology, chemical composition, and mineralogy of plume-front colloids collected from the peak turbidity fraction are presented in Figures 3 (21 °C) and 4 (70 °C). Figure 3a shows the typical morphology of colloids formed at ambient temperature. This image indicates that the nominally 1 μm size is dominant in terms of mass and the nominally 0.1 μm size is dominant in terms of numbers of particles. For the sample shown in Figure 3a, powder XRD analyses identified

only calcite (Figure 3b), and the fine particles presented in Figure 3a were amorphous. EDX analyses indicated that the individual larger particles in Figure 3a consist of Ca–O–C (not shown), while the total area (10 μm^2) analyses yielded Ca–Al–Mg–O–C (Figure 3c). In another 21 °C plume-front sample, XRD patterns matched the spectra of calcium magnesium aluminum oxide carbonate hydrate [(Ca,Mg)O]₃·Al₂O₃·CaCO₃·11H₂O} (not shown). In contrast, the 70 °C plume-front colloids are very different in most respects from the 21 °C colloids, except that Ca is also a major element. Figure 4 panels a and b show the overall morphology of the particles and the area-averaged chemical composition, respectively. The major elements of these colloids are Si–Ca–O–C. Figure 4a shows two types of morphologies: fine particles (Figure 4c) and larger hollow spheroids (Figure 4d). EDX data indicate that the fine particles are CaCO₃-rich and also contain small amounts of Al, Mg, Na, Si, and S. The larger spheroids are predominantly Si–O but also contain Na and Mg. Like the ambient temperature colloids, XRD analyses from the 70 °C samples identified only calcite. The Si-rich spheroidal aggregates were amorphous in TEM electron microdiffraction analyses as well.

We also studied the morphology, chemical composition, and mineralogy of particles collected from the plume body and leaching stage effluents. We found that, within the plume body effluents, cancrinite [Na₈(Al₆Si₆O₂₄)(NO₃)₂·4H₂O] and zeolite [Na₂Al₂Si_{1.85}O_{7.7}·5.1H₂O] were the major mineral phases under both temperatures, with larger quantity of precipitates under 70 °C. This is because of higher Si concentrations from higher silicate dissolution rates at 70 °C. Particles collected from the leaching stages under the two temperatures were very different. At 70 °C, the same cancrinite and zeolite found in the plume body were also identified in leaching stage. However, at 21 °C, bayerite and calcite were the major mineral phases. Unlike the plume-front colloids (dominantly amorphous/poorly crystallized and submicrometer sizes), these precipitates were well crystallized and larger (up to about 5 μm). The majority of these formed after the effluents were collected. Details of these results are not presented because this paper is focused on the plume fronts.

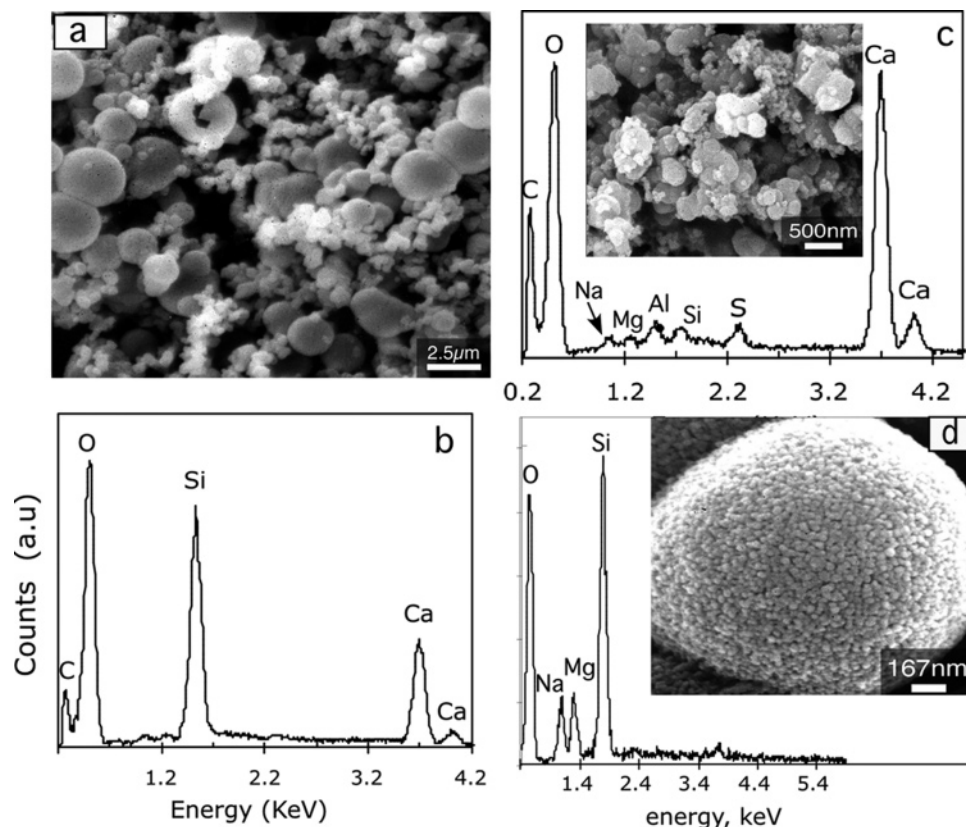


FIGURE 4. Plume-front colloids at 70 °C. (a) SEM image of a bulk sample. (b) EDX of the bulk sample. (c) Fine particles contained predominantly Ca–C–O, and calcite was identified in this sample. (d) Spheres are aggregates of fine amorphous Si–O–(Na,Mg) particles.

TABLE 5. Geochemical Reactions within Plume Front

possible plume-front reactions	eq	log <i>K</i> (25 °C)
$\text{Na}^+ + \frac{1}{2}\text{Ca/Mg-X} = \text{Na-X} + \frac{1}{2}\text{Ca}^{2+}/\text{Mg}^{2+}$	(1)	–0.4 ^a /–0.3 ^a
$\text{Ca}^{2+} + \text{HCO}_3^- = \text{CaCO}_3 \text{ (calcite)} + \text{H}^+$	(2)	–1.85 ^b
$\text{Ca}^{2+} + \text{Mg}^{2+} + 2\text{HCO}_3^- = \text{CaMg}(\text{CO}_3)_2 \text{ (dolomite)} + 2\text{H}^+$	(3)	–3.66 ^b
$\text{Mg}^{2+} + 2\text{H}_2\text{O} = \text{Mg}(\text{OH})_2 \text{ (brucite)} + 2\text{H}^+$	(4)	–16.79 ^b
$2\text{Al}(\text{OH})_4^- + 4(\text{Ca,Mg})^{2+} + \text{CO}_3^{2-} + 9\text{H}_2\text{O} = [(\text{Ca,Mg})\text{O}]_3\cdot\text{Al}_2\text{O}_3\cdot\text{CaCO}_3\cdot 11\text{H}_2\text{O} + 4\text{H}^+$	(5)	40.35 ^c
$\text{Al}(\text{OH})_4^- = \text{Al}(\text{OH})_3 \text{ (bayerite)} + \text{OH}^-$	(6)	8.36 ^b
$\text{H}_2\text{SiO}_4^{2-} + 2\text{Na}^+ = \text{Na}_2\text{SiO}_3 \text{ (sodium metasilicate)} + \text{H}_2\text{O}$	(7)	0.81 ^d

^a Ref 24. ^b Ref 25. ^c Ref 14. ^d Ref 26.

Particle Size. Although the turbidity data presented in Figures 1 and 2b were measured as soon as an effluent fraction was collected, the SEM and XRD samples were prepared about 4 weeks later. This time lapse and equilibration with ambient temperature contributed to particle formation in effluents from the 70 °C columns. Thus, the amorphous, hollow silicate spheres (Figure 4a,d) likely grew and aggregated after collection of the effluent. The particle size measurements obtained on freshly collected effluent suspensions are presented in Figure 5, only for the ambient temperature plume-front colloids. Because these measurements must be done at ambient temperature, samples collected from the 70 °C column would be influenced by postsampling precipitation. The particle size analyses are from duplicate samples and are expressed in terms of relative masses, relative numbers, and relative surface areas, assuming a density of 2.7 g cm^{–3} and spherical shape. In terms of relative mass (Figure 5a), over 90% of the mass is associated with colloids in the size range of 0.5–2.3 μm, with the highest contribution by about 1.0 μm sized particles. In terms of number (Figure 5b), the large contribution is by small particles in the size range of 30–100 nm (30 nm is the resolution limit of the CPS). In terms of surface area (Figure 5c), over 90% of the

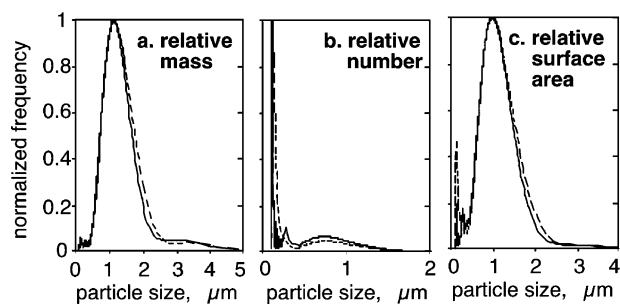


FIGURE 5. Size distribution of plume-front colloids at 21 °C, presented in terms of (a) relative particle mass, (b) relative particle number, and (c) relative surface area. The particle density was assumed to be 2.7 g cm^{–3}, and the shape was assumed to be spherical. The two sets of curves are duplicate samples of the plume-front fraction.

total surface area is associated with colloids in the size range of 0.3–2.0 μm. It is interesting to note that although the <0.1 μm fraction is dominant in terms of particle numbers, the nominally 1 μm fraction is dominant in terms of not only mass but also surface area.

Mechanisms Responsible for Colloid Formation at the Plume Front. The coupled transport and geochemical reactions during the tank leakage process are complex, and our understanding of these reactions and processes is still incomplete. Based on our limited experimental data, the likely geochemical reactions controlling colloid formation at the plume front are summarized in Table 5. Cation exchange (eq 1), with Na^+ replacing Ca^{2+} and Mg^{2+} from the cation exchange sites of the sediment, was the key process contributing to this plume-front phenomenon. The cation-exchange reaction was rapid and complete. At the same time, transport kept the exchanged divalent cations within the moving front of the waste plume, where $\text{Ca}^{2+}/\text{Mg}^{2+}$ accumulated to elevated levels. The measured $\text{Ca}^{2+}/\text{Mg}^{2+}$ concentrations in the plume front of the 250-mm column were as high as 0.61 M/0.14 M at 21 °C and 0.38 M/0.07 M at 70 °C (Figure 2d,e) and 1.2 M/0.4 M for a 2-m tall column at 70 °C (13). Supersaturation of these divalent cations with respect to a number of $\text{Ca}^{2+}/\text{Mg}^{2+}$ -bearing minerals caused their rapid precipitation as well as precipitation of amorphous phases. Calcite (eq 2) was identified at both temperatures. Although dolomite and brucite (eqs 3 and 4) were not detected, we speculate that they could have formed in minor quantities. Calcium magnesium aluminum oxide carbonate hydrate and bayerite were identified (eqs 5 and 6); however, in most of the analyses, amorphous phases containing Ca–Mg–Al–O–C were common in ambient temperature samples. Temperature is an important factor controlling the chemical composition of the colloid phases in these caustic plumes, primarily through affecting silicate dissolution rate, and consequently aqueous Si concentration, pH, and in turn aqueous Al concentration. At ambient temperature, the concentration of the dissolved Si was not high enough to coprecipitate with much of the Al (0.8 M) initially contained in the TWS, and the solution pH was high enough to keep most of this Al in the solution as $\text{Al}(\text{OH})_4^-$. Therefore, we saw high concentrations of Al-rich particles at the plume front only at the ambient temperature. Because of the extreme alkalinity of the initial TWS (pH ≥ 14), dissolution of primary silicates in the sediment was significant, especially at elevated temperatures. Some of the dissolved Si was advected with the plume front, and after encountering the reduced pH at the plume front it precipitated, forming the colloidal Si phase (eq 7). Amorphous sodium silicate (eq 7) was abundant within the 70 °C samples. Although cancrinite/zeolites precipitates were identified as the major secondary phases in plume bodies at both temperatures, they were not detected at the plume front.

Implications for Radionuclide Transport. This study revealed a striking phenomenon during leakage of a highly saline–caustic waste solution into sediments: colloid formation within a narrow zone at the moving plume front. Transport of suspended colloids was thousands of times higher in this narrow plume front than in the plume body and the leaching stage. The newly formed plume-front colloids were primarily in the size range of 0.3–2.0 μm , predominantly 1.0 μm , and poorly crystallized. The major chemical compositions of these colloids were Ca–Mg–Al–C–O at ambient temperature and Ca–Mg–Si–C–O at elevated temperature. Rapid and completed cation exchange of Na^+ replacing Ca^{2+} and Mg^{2+} from the sediments, coupled with flow of the infiltrating waste solution, enriched these divalent cations within the moving plume front. Subsequent precipitation of colloids containing these divalent cations released protons and reduced the plume-front pH to as low as neutral. This substantially reduced pH in turn promoted precipitation of Al- and Si-bearing minerals and amorphous phases. This plume-front colloid generation phenomenon can occur under a wide range of conditions, with the only necessary condition being that of high Na^+ concentration in

the waste plume. Because of the large amounts of particles generated at the front, this phenomenon could significantly affect the fate and transport of the contaminant trace elements contained within the waste plume. Depending on the chemical properties of the individual trace element, radionuclide or toxic metal, it can coprecipitate with or adsorb onto the plume-front colloids. For example, U sorption and coprecipitation on calcite (27) may be important in some contaminated sediments at the Hanford Site because high levels of U are contained in some tank wastes and because calcite is the major plume-front colloid phase.

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Note Added after ASAP Posting

This paper was released ASAP on October 7, 2004, with an error in the 9th line of the last paragraph of the paper. The corrected version was posted on October 14, 2004.

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